

## Supporting Information

# Enhanced electron transfer in Fe–N–C catalysts for nitrobenzene reduction: From electrodes to functional materials

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**Fig. S4.** Electrochemical performance and stability of different catalysts. (a) Cyclic voltammetry (CV) curves, and (b) cycling stability of Fe–N–C, NPCs, and GR electrodes. (c) Iron leaching was

observed during 1000 linear sweep voltammetry (LSV) cycles. (d) XRD patterns of the post-cycling Fe–N–C electrode. (Scan rate=50 mV/s, [Na<sub>2</sub>SO<sub>4</sub>]=50 mM, [NB]=100 mg/L).

**Fig. S5.** Adsorption of nitrobenzene (NB) by electrodes prepared from active carbons (ACs), nitrogen-doped porous carbons (NPCs), and Fe–N–C. (pH=6.1, [NB]<sub>0</sub>=100 mg/L, [Na<sub>2</sub>SO<sub>4</sub>]<sub>0</sub>=50 mM, T=298 K).

**Fig. S6.** Effect of iron content in Fe–N–C electrodes on the electrocatalytic performance of nitrobenzene (NB). (inset: pseudo-first-order kinetic fitting). Experimental conditions: pH=6.1, [NB]<sub>0</sub>=100 mg/L, [Na<sub>2</sub>SO<sub>4</sub>]<sub>0</sub>=50 mM, T=298 K.

**Fig. S7.** Structural and compositional characterization of FeNCs-500 catalyst. (a) XRD patterns, and (b) high-resolution N 1s XPS spectra.

**Fig. S8.** Electrochemical reduction performance of NB using different electrocatalysts at varying pH conditions. (a) NB removal efficiency and (b) kinetics with GR electrode; (c) NB removal efficiency and (d) kinetics with NPCs electrode. (T=298 K, [NB]<sub>0</sub>=100 mg/L, [Na<sub>2</sub>SO<sub>4</sub>]<sub>0</sub>=50 mM, potential=-1.0 V vs. Ag/AgCl).

**Fig. S9.** (a) Reduction efficiency of NB and (b) its pseudo-first-order rate constant ( $k_{\text{obs}}$ ) ratio on the Fe–N–C cathode with different *t*-BuOH concentrations (Potential=-1.0 V vs. Ag/AgCl, T=298 K, [NB]<sub>0</sub>=100 mg/L, [Na<sub>2</sub>SO<sub>4</sub>]<sub>0</sub>=50 mM, electrolysis time=180 min).

**Fig. S10.** Morphological and elemental characterization of Fe–N–C/ZVI. (a and b) FE-SEM images, (c) EDS elemental distribution, and (d) EDS elemental mapping images of Fe–N–C/ZVI. The inset of (c) shows the chemical composition of the samples.

**Fig. S11.** XRD patterns of the Fe–N–C/ZVI composite as well as Fe (PDF#85-1410, yellow vertical line) and Fe<sub>2</sub>O<sub>3</sub> (PDF#89-0597, black vertical line) as references.

**Fig. S12.** High-resolution XPS spectra of (a) survey scan, (b) Fe 2p, (c) O 1s, and (d) C 1s for the Fe–N–C/ZVI composite.

**Fig. S13.** Parametric optimization of Fe–N–C/ZVI composites for nitrobenzene (NB) degradation. (a) Effect of mZVI/Fe–N–C mass ratio ([dosage]=1.0 g/L, ball-milling time: 4 h), (b) Ball-milling duration ([dosage]=1.0 g/L, mass ratio=1:20), and (c) feeding dosage (ball-milling time: 4 h, mass ratio=1:20). (pH=6.1, [NB]<sub>0</sub>=100 mg/L, 25 °C)

**Fig. S14.** Effects of common substances on the NB removal efficiencies by Fe–N–C/ZVI and mZVI: (a) chloride, (b) sulfate, (c) bicarbonate, (d) humic acid ([NB]<sub>0</sub>=100 mg/L, T=298 K, pH=5.0, reaction time=120 min).

**Fig. S15.** Comparative evaluation of NB removal efficiency and kinetic behavior under varying pH conditions. Subfigure pairs (a-b) pH=3, (c-d) pH=5, (e-f) pH=7, and (g-h) pH=9 show temporal removal profiles (left panels) and corresponding kinetic fittings (right panels), respectively. ([NB]<sub>0</sub>=100 mg/L, T=298 K).

**Fig. S16.** The cyclic fixed-bed decontamination performances of Fe–N–C, mZVI, and Fe–N–C/ZVI toward (a) synthesized wastewater and (b) real wastewater collected from an industrial park (T=298 K, flow rate=6 BV/h).

**Fig. S17.** The iron ion concentration was detected in the effluent from continuous-flow degradation experiments. ([NB]<sub>0</sub>=60 mg/L, T=298 K, flow rate=6 BV/h).

**Scheme S1.** Proposed pathway for NB electrochemical reduction by Fe–N–C electrode.

## Supplementary Text, Figures and Tables

### Text S1. Electrochemical Analysis

#### (1) Preparation of coated glassy carbon electrode

A catalyst-coated glassy carbon electrode was prepared by dipping catalytic ink on the glassy carbon electrode (GCE). The catalyst powder (10 mg) was dispersed in a mixed solution containing 780  $\mu\text{L}$  of water, 200  $\mu\text{L}$  of ethanol, and 20  $\mu\text{L}$  of Nafion (5 wt%) by 30-min ultrasonication. A volume of 10  $\mu\text{L}$  of the resulting catalyst ink was drop-cast onto the glassy carbon (GC) electrode surface and then dried at room temperature. Before use, the GC electrode ( $\Phi = 5$  mm) was polished with 50 nm  $\text{Al}_2\text{O}_3$  powder and then washed with deionized water.

#### (2) Electrochemical active surface area (ECSA) measurement

Various catalysts' electrochemical active surface area (ECSA) was determined by measuring the capacitive current associated with double-layer charging from the scan-rate dependence of CV curves. ECSA measurements were performed in a 0.5 M  $\text{Na}_2\text{SO}_4$  electrolyte using a three-electrode setup, with a Pt plate as the counter electrode, an Ag/AgCl electrode as the reference, and a catalyst-coated GCE as the working electrode. The CV measurements were conducted within a potential window of 0.3-0.4 V vs. Ag/AgCl, and the scan rates employed were 20, 40, 60, 80, and 100  $\text{mV s}^{-1}$ . The double-layer capacitance ( $C_{dl}$ ) of each catalyst was obtained by plotting the difference between the anodic and cathodic current densities at 0.35 V vs. Ag/AgCl against the scan rate. The slope of this linear plot corresponds to twice the double-layer capacitance ( $C_{dl}$ ), which serves as a proxy for the ECSA.

### **Text S2. Fabrication of self-supported Fe–N–C electrode**

Self-supported PTFE membrane electrodes were fabricated using a casting method for the reduction of nitrobenzene (NB). The electrode composition comprised the as-synthesized catalyst as the active material, carbon black (CB) as the conductive agent, and polytetrafluoroethylene (PTFE, 10 wt% aqueous dispersion) as the binder. These components were homogenized in a mass ratio of 8:1:1 (catalyst: CB: 10 wt% PTFE) under sequential ultrasonication (30 min) and mechanical stirring. Anhydrous ethanol was incrementally added during stirring to promote fibrillation, resulting in a cohesive, agglomerated paste. The paste was then uniformly roll-pressed onto a titanium (Ti) mesh current collector using a heated roller system (120 °C) to form the PTFE membrane electrode. The resulting electrodes exhibited an active material loading density of 4.7~5.0 mg·cm<sup>-2</sup>.

### **Text S3. Nitrobenzene reduction by Fe–N–C /ZVI composite**

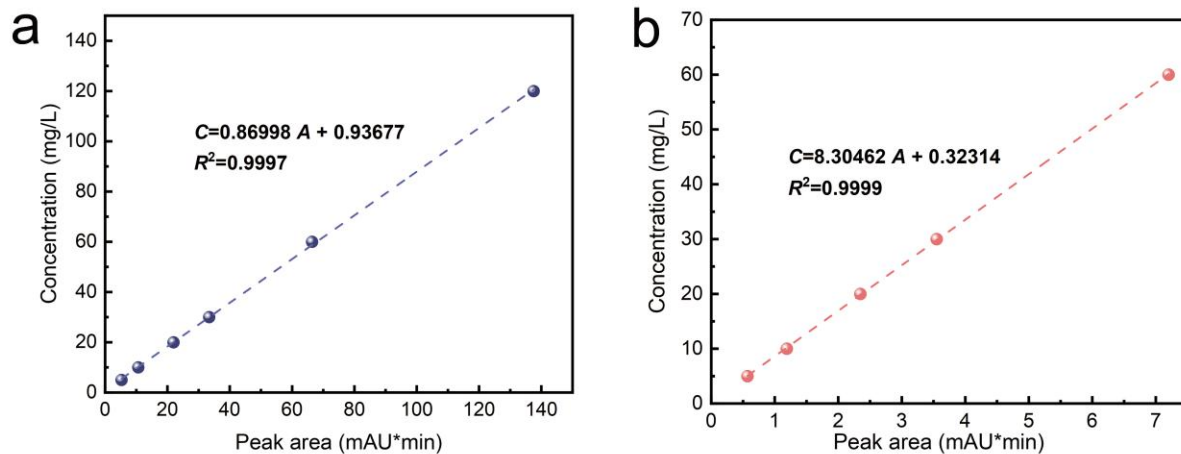
Batch reduction experiments were conducted in a 60 mL amber glass reactor under ambient aerobic conditions. The reaction system contained 50 mL of nitrobenzene (NB) solution (100 mg·L<sup>-1</sup> prepared by diluting a stock solution (1 g·L<sup>-1</sup>), with initial pH maintained at 6.1±0.1 without buffer addition. For each trial, 0.05 g of zero-valent iron (ZVI) or Fe–N–C/ZVI composite was introduced into the reactor under constant magnetic stirring (600 rpm). Aliquots (0.5 mL) were periodically withdrawn, immediately filtered through 0.45 µm polyethersulfone (PES) membranes, and stored at 4 °C for subsequent analytical characterization. The pH of the solution was adjusted to predetermined values (3.0-9.0) to evaluate the effects of pH using 1 M HCl or NaOH solutions. All experiments were performed in triplicate with parallel controls at 25°C.

#### **Text S4. Continuous-flow degradation experiment**

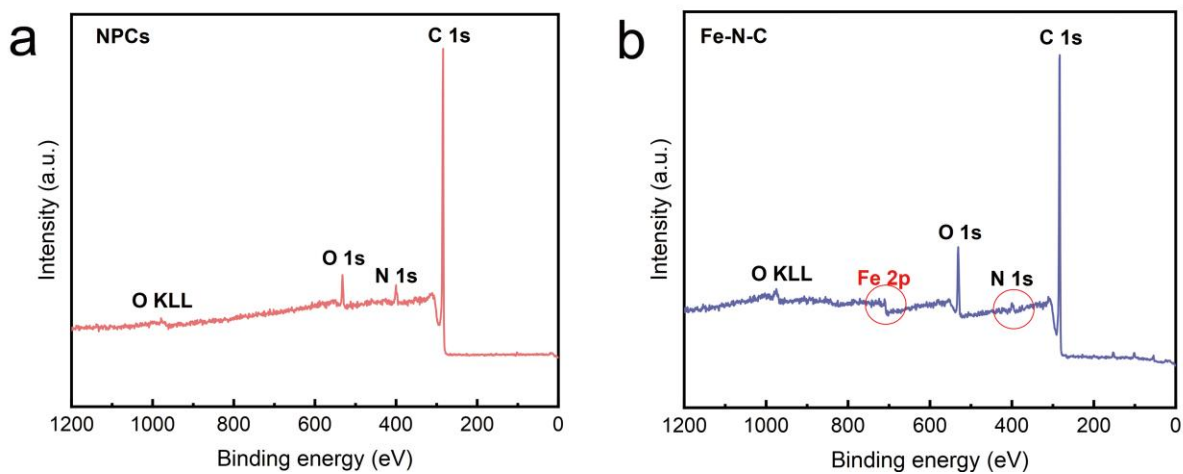
Fixed-bed degradation experiments were conducted in a glass column (300 mm length, 20 mm diameter) filled with 5 mL of wet Fe–N–C/ZVI. Both synthetic and real wastewater, collected from an industrial park containing approximately 60 mg/L of NB and other competing anions, were fed into the columns using a constant-flow peristaltic pump. The empty bed contact time (EBCT) for these experiments was set to 10 minutes, corresponding to a flow rate of 6 bed volumes per hour (BV/h). After the degradation process, the effluent was collected and analyzed for further evaluation.

#### **Text S5. Quantification of nitrobenzene (NB) and aniline (AN) by HPLC.**

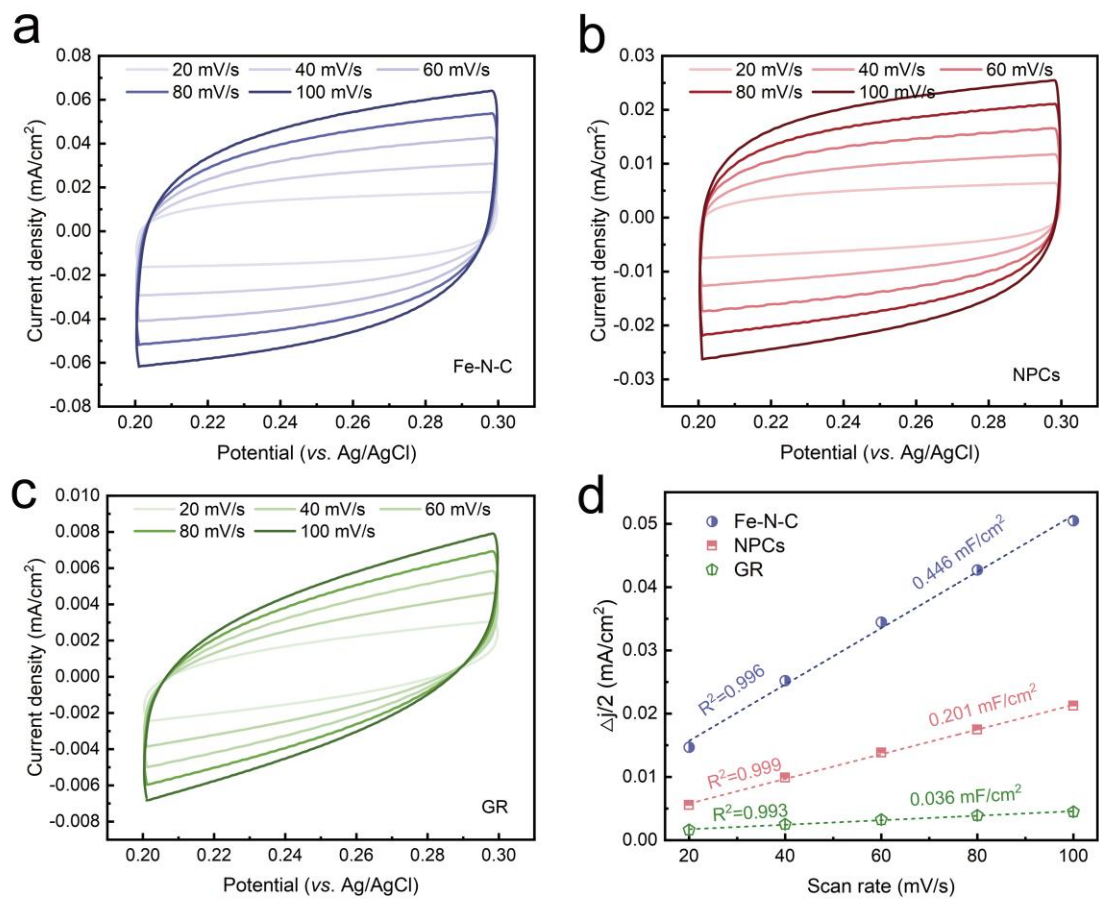
Residual nitrobenzene (NB) and aniline (AN) concentrations in treated wastewater were quantified using high-performance liquid chromatography (HPLC) under the following operational parameters: Column: C18 reverse-phase column (250 mm × 4.6 mm, 5 μm particle size); Detection: UV absorbance at 254 nm; Mobile phase: Methanol/water (60:40, v/v); Flow rate: 1.00 mL·min<sup>-1</sup>; Column temperature: 35 °C; Injection volume: 30 μL. Chromatographic separation yielded distinct retention times of 3.2 min for AN and 7.8 min for NB. Calibration curves (Fig. S1) were established using serially diluted standard solutions (5-120 μg·mL<sup>-1</sup>), demonstrating excellent linearity: AN:  $C=8.30462 A+ 0.32314$  ( $R^2=0.9999$ ), and NB:  $C=0.86998 A + 0.93677$  ( $R^2=0.9997$ ). Where  $C$  represents concentration (μg·mL<sup>-1</sup>), and  $A$  corresponds to the chromatographic peak area. All samples were filtered through 0.45 μm PES membranes before injection.



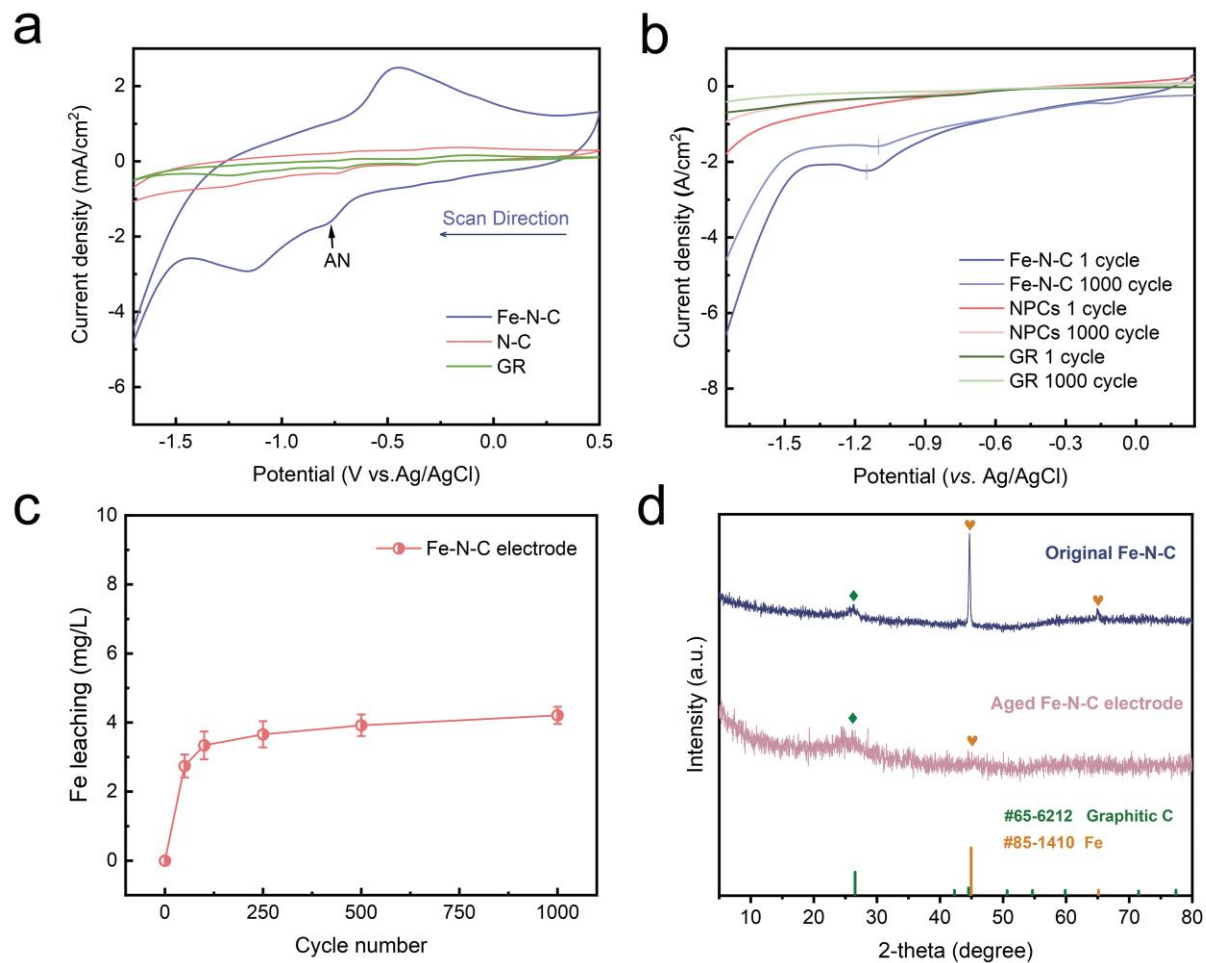
**Fig. S1.** Calibration curves for the quantification of (a) nitrobenzene (NB) and (b) aniline (AN) by HPLC.



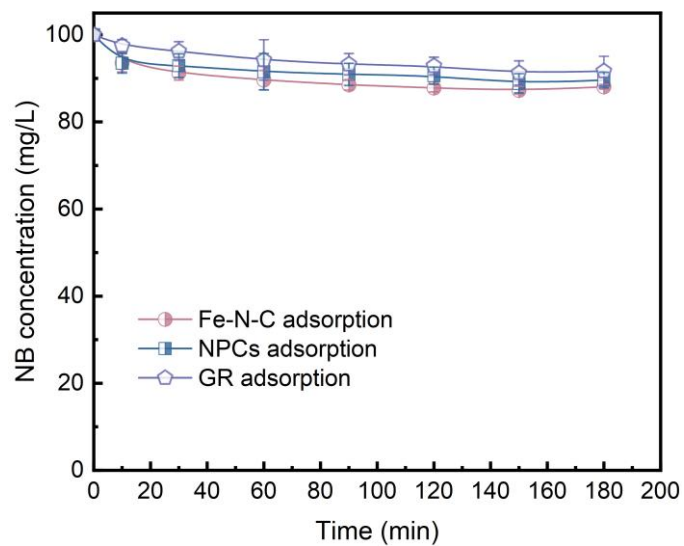
**Fig. S2.** Survey XPS spectra of (a) NPCs and (b) Fe-N-C materials.



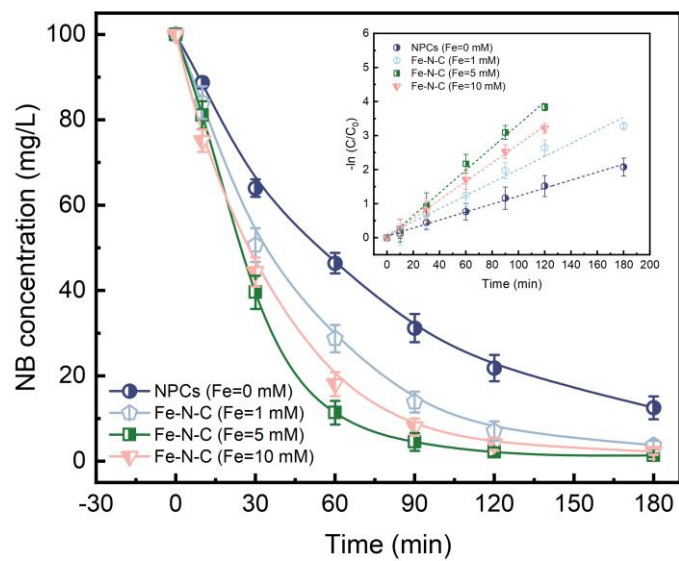
**Fig. S3.** CV curves of (a) Fe–N–C, (b) nitrogen-doped porous carbons (NPCs), (c) graphite (GR), and (d) ECSA measurements of catalysts based on the double-layer capacitance ( $C_{dl}$ ) over a range of scan rates. (pH=7.0,  $[\text{Na}_2\text{SO}_4]_0=0.5 \text{ M}$ ,  $T=298 \text{ K}$ ).



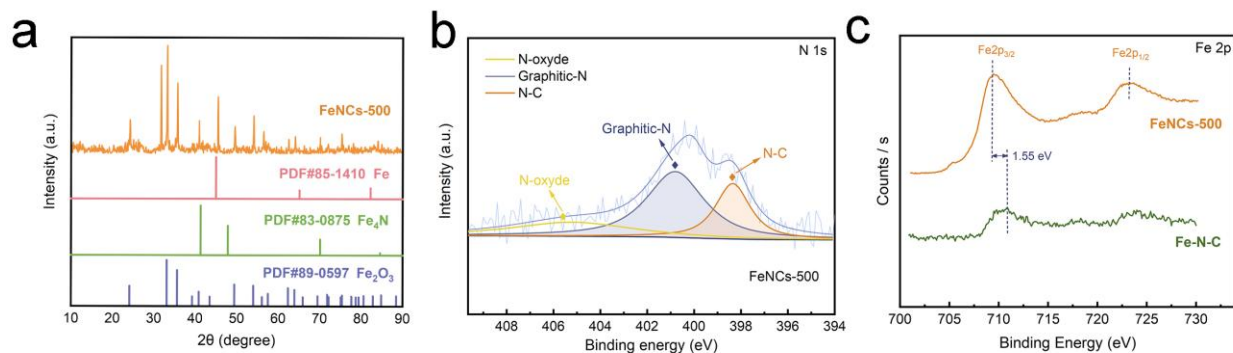
**Fig. S4.** Electrochemical performance and stability of different catalysts. (a) Cyclic voltammetry (CV) curves, and (b) cycling stability of Fe–N–C, NPCs, and GR electrodes. (c) Iron leaching observed during 1000 linear sweep voltammetry (LSV) cycles. (d) XRD patterns of the post-cycling Fe–N–C electrode. (Scan rate=50 mV/s, [Na<sub>2</sub>SO<sub>4</sub>]<sub>0</sub>=50 mM, [NB]<sub>0</sub>=100 mg/L).



**Fig. S5.** Adsorption of nitrobenzene (NB) by electrodes prepared from active carbons (ACs), nitrogen-doped porous carbons (NPCs), and Fe-N-C. (pH=6.1,  $[NB]_0=100$  mg/L,  $[Na_2SO_4]_0=50$  mM, T=298 K).

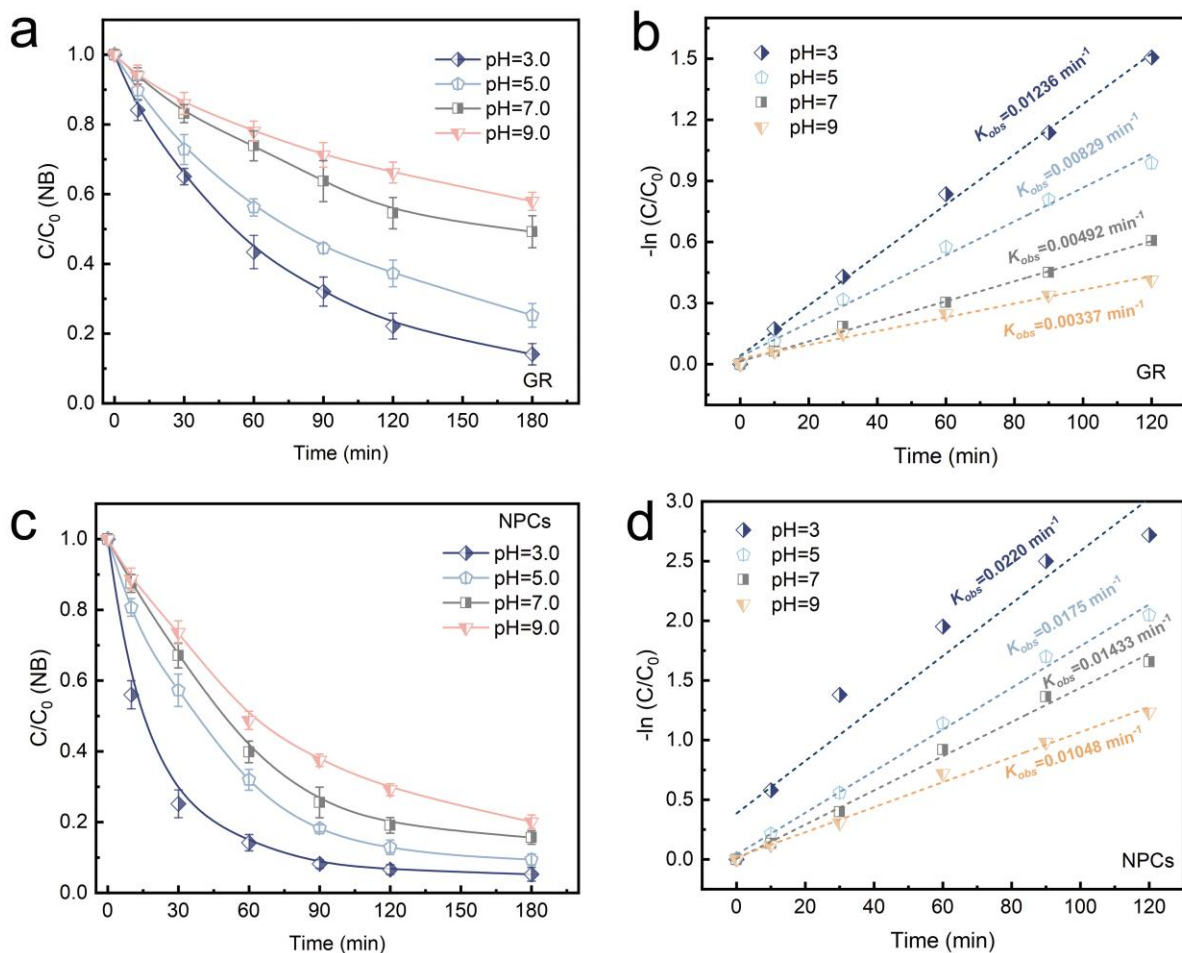


**Fig. S6.** Effect of iron content in Fe–N–C electrodes on the electrocatalytic performance of nitrobenzene (NB). (inset: pseudo-first-order kinetic fitting). Experimental conditions: pH=6.1,  $[NB]_0=100$  mg/L,  $[Na_2SO_4]_0=50$  mM,  $T=298$  K.

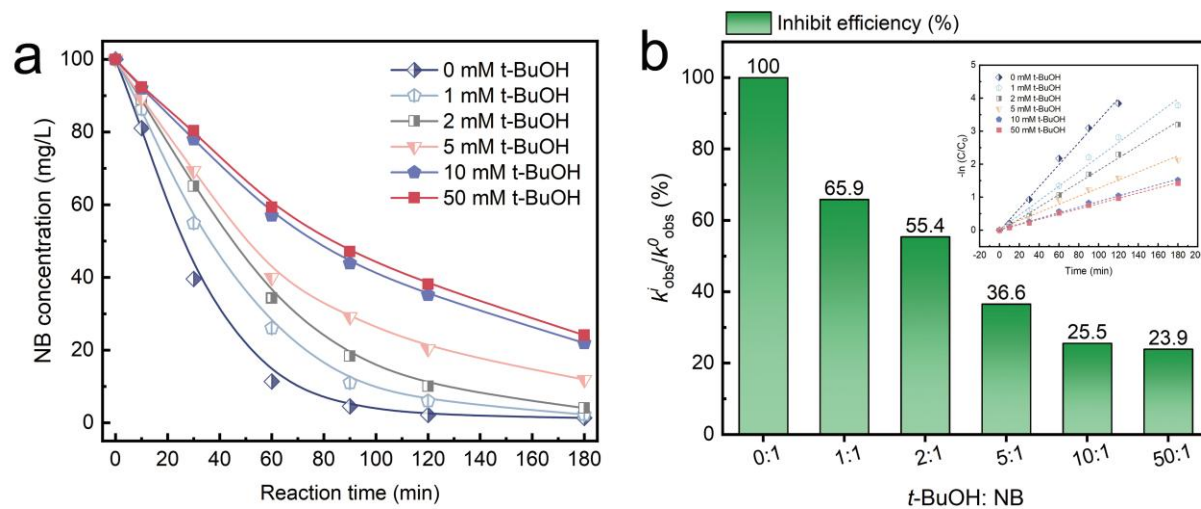


**Fig. S7.** Structural and compositional characterization of FeNCs-500 catalyst. (a) XRD patterns, and (b) high-resolution N 1s XPS spectra. (c) The Fe 2p narrow scan spectra of FeNCs-500 and Fe–N–C catalysts.

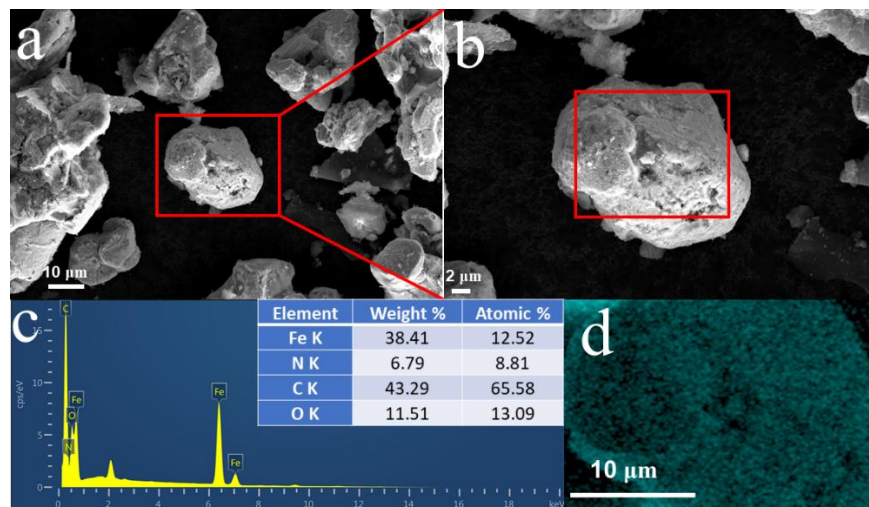
The synthesis protocol for FeNCs-500 is detailed in Section 2.3. XRD analysis revealed the predominant presence of iron trioxide crystalline phases in the catalyst matrix. Deconvolution of the N 1s XPS spectrum demonstrated that the low calcination temperature (500°C) resulted in incomplete nitrogen coordination, as evidenced by the absence of characteristic Fe–N<sub>x</sub> bonding configurations typically observed in higher-temperature (800°C) pyrolyzed catalysts (Fe–N–C).



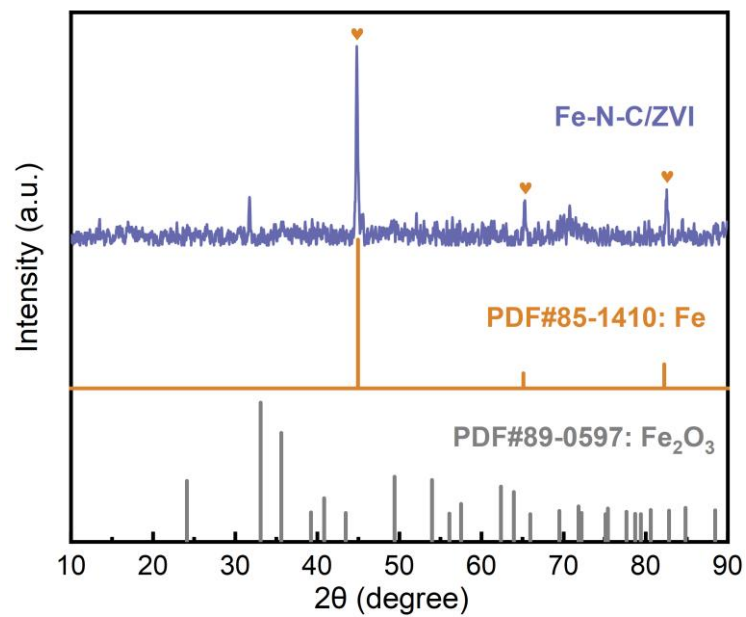
**Fig. S8.** Electrochemical reduction performance of NB using different electrocatalysts at varying pH conditions. (a) NB removal efficiency and (b) kinetics with GR electrode; (c) NB removal efficiency and (d) kinetics with NPCs electrode. ( $T=298 \text{ K}$ ,  $[\text{NB}]_0=100 \text{ mg/L}$ ,  $[\text{Na}_2\text{SO}_4]_0=50 \text{ mM}$ , potential= $-1.0 \text{ V vs. Ag/AgCl}$ ).



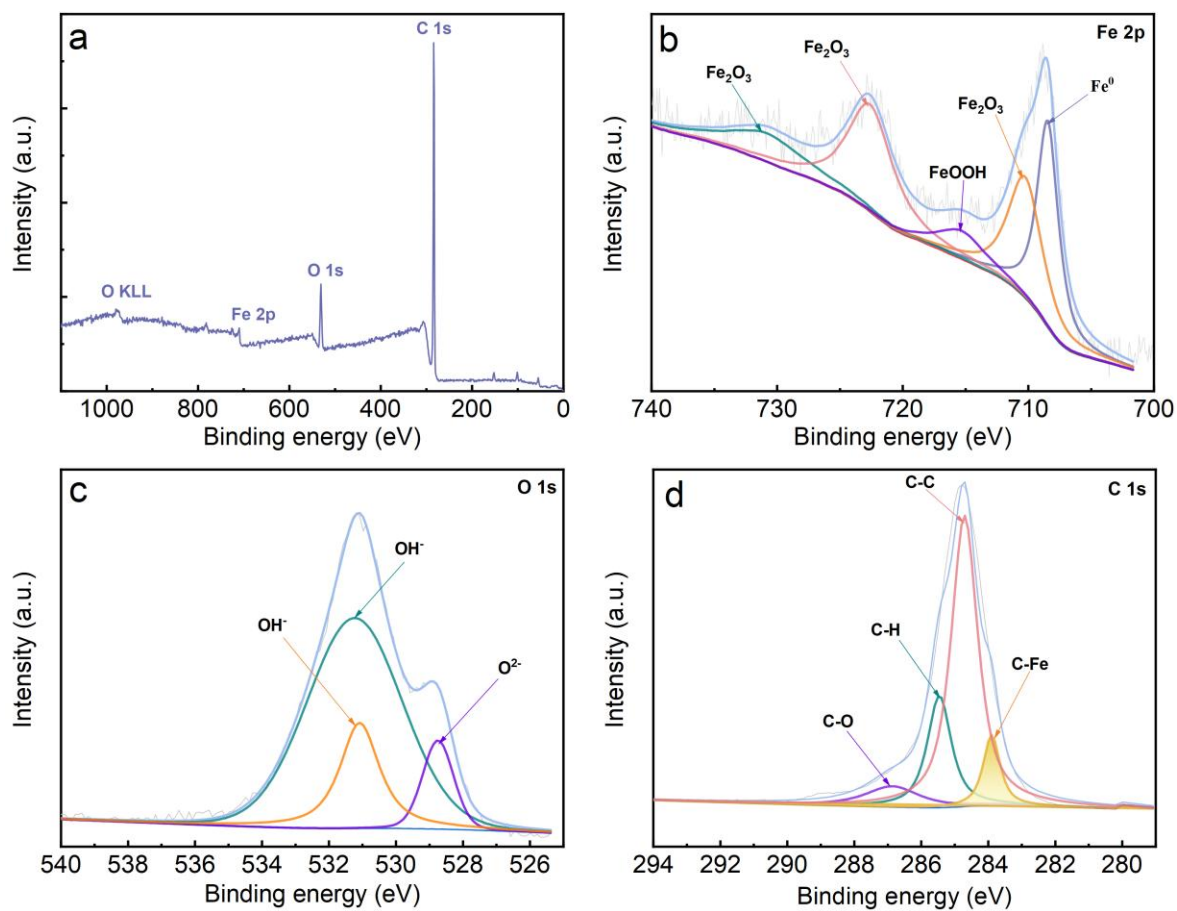
**Fig. S9.** (a) Reduction efficiency of NB and (b) its pseudo-first-order rate constant ( $k_{obs}$ ) ratio on the Fe–N–C cathode with different *t*-BuOH concentrations (Potential=-1.0 V vs. Ag/AgCl, T=298 K, [NB]<sub>0</sub>=100 mg/L, [Na<sub>2</sub>SO<sub>4</sub>]<sub>0</sub>=50 mM, electrolysis time=180 min).



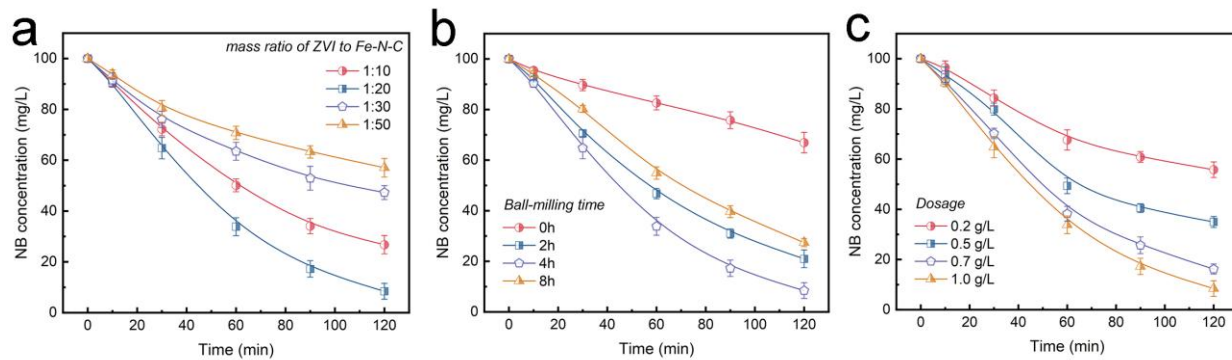
**Fig. S10.** Morphological and elemental characterization of Fe–N–C/ZVI. (a and b) FE-SEM images, (c) EDS elemental distribution, and (d) EDS elemental mapping images of Fe–N–C/ZVI. The inset of (c) shows the chemical composition of the samples.



**Fig. S11.** XRD patterns of the Fe-N-C/ZVI composite as well as Fe (PDF#85-1410, yellow vertical line) and Fe<sub>2</sub>O<sub>3</sub> (PDF#89-0597, black vertical line) as references.

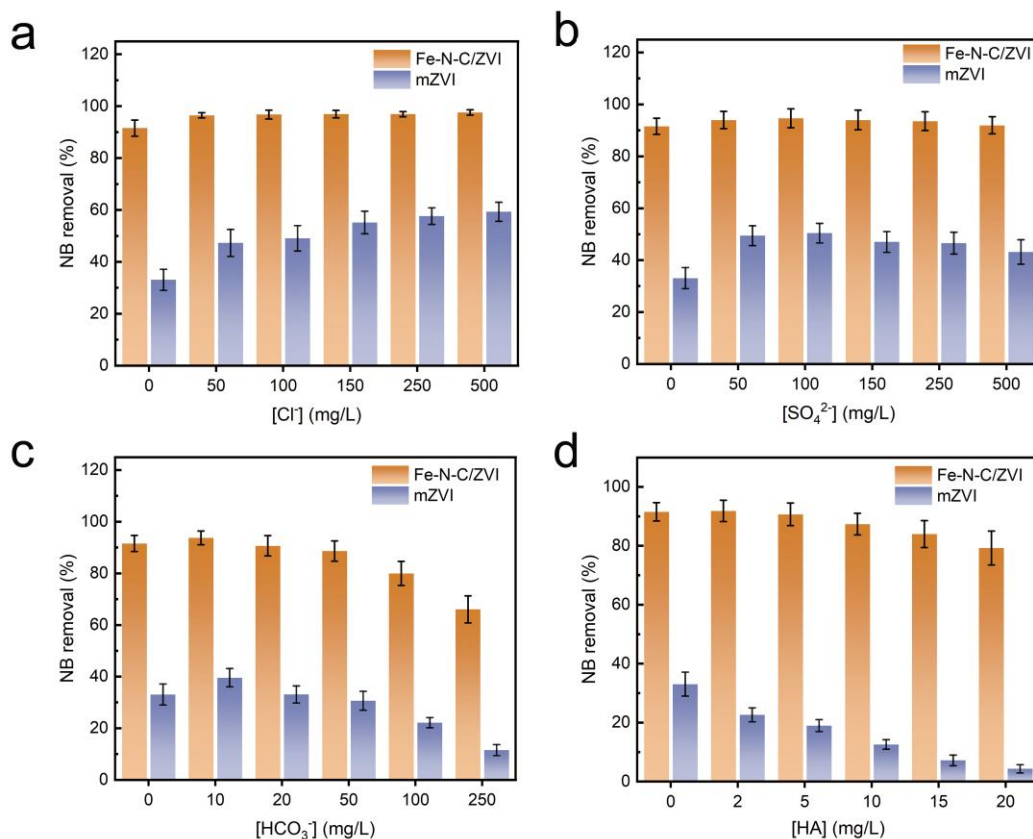


**Fig. S12.** High-resolution XPS spectra of (a) survey scan, (b) Fe 2p, (c) O 1s, and (d) C 1s for the Fe-N-C/ZVI composite.



**Fig. S13.** Parametric optimization of Fe–N–C/ZVI composites for nitrobenzene (NB) degradation.

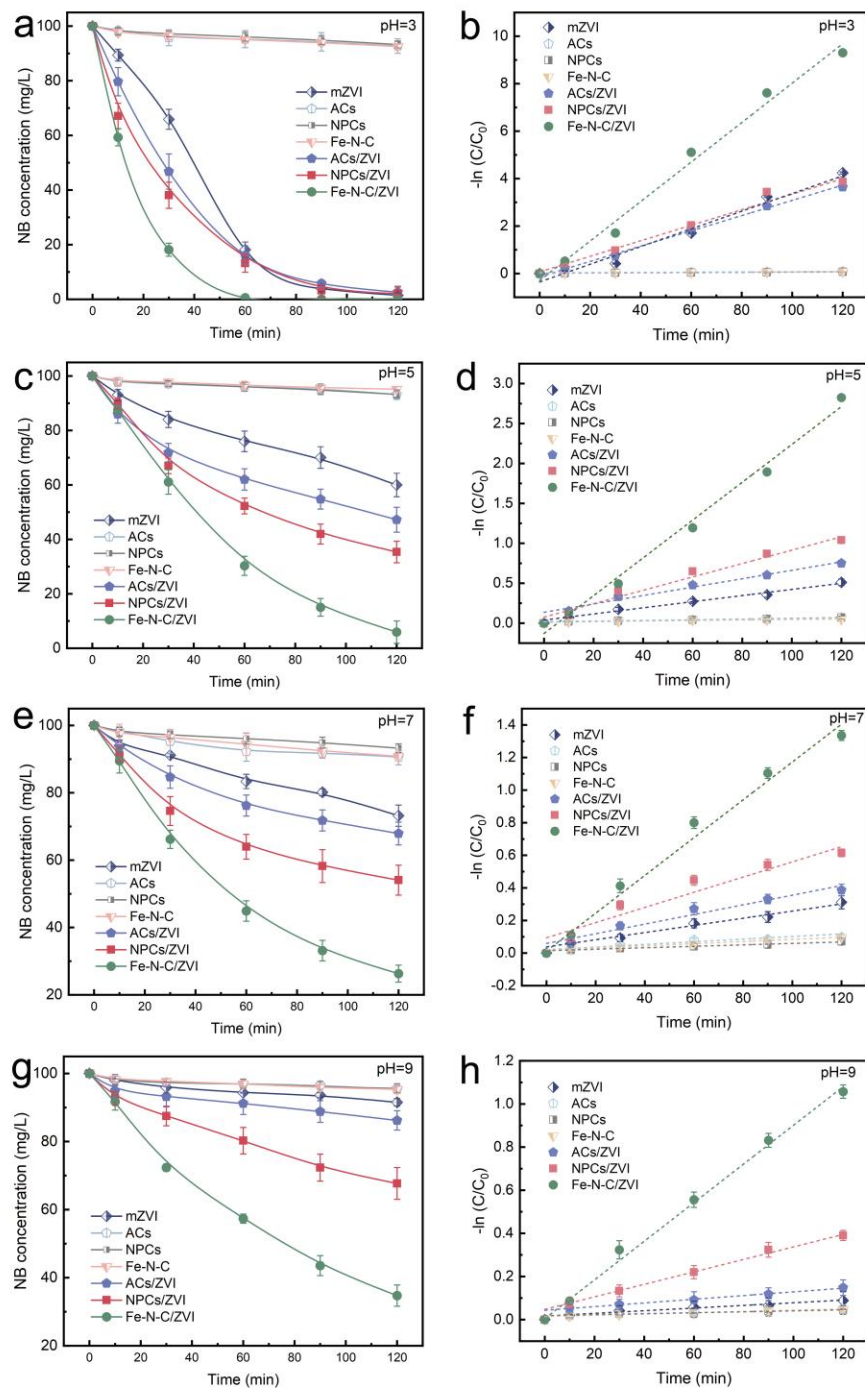
(a) Effect of mZVI/Fe–N–C mass ratio ([dosage]=1.0 g/L, ball-milling time: 4 h), (b) Ball-milling duration ([dosage]=1.0 g/L, mass ratio=1:20), and (c) feeding dosage (ball-milling time: 4 h, mass ratio=1:20). (pH=6.1, [NB]<sub>0</sub>=100 mg/L, 25 °C)



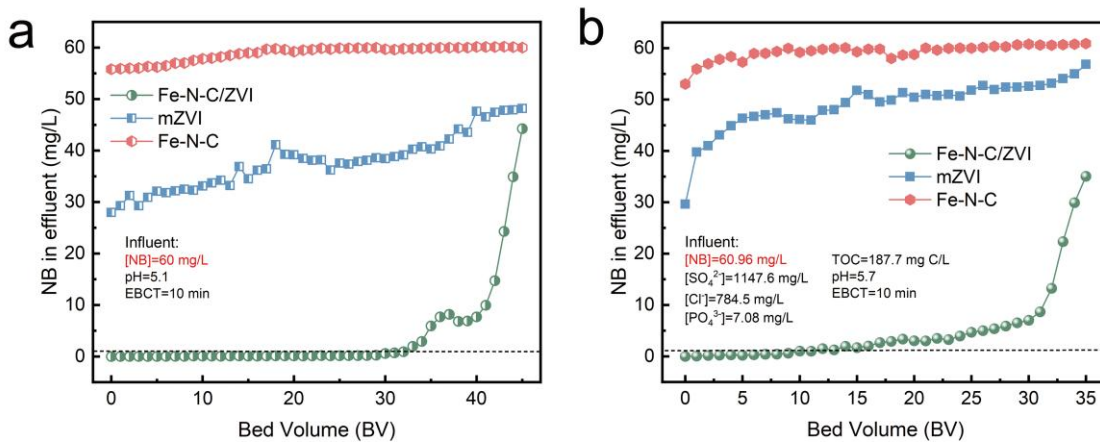
**Fig. S14.** Effects of common substances on the NB removal efficiencies by Fe–N–C/ZVI and mZVI: (a) chloride, (b) sulfate, (c) bicarbonate, (d) humic acid ([NB]<sub>0</sub>=100 mg/L, T=298 K, pH=5.0, reaction time=120 min).

**Note:** As shown in Fig. S14a and S14b, the introduction of SO<sub>4</sub><sup>2-</sup> and Cl<sup>-</sup> slightly enhanced nitrobenzene (NB) removal, primarily due to their ability to disrupt the passivating oxides on the iron surface, thereby regenerating the active sites (Huang and Zhang, 2006; Wang et al., 2006). Fig. S14c indicates that low concentrations of bicarbonate (<50 mg/L) had negligible effects on both mZVI and Fe–N–C/ZVI. However, when the bicarbonate concentration was increased to 250 mg/L, the removal efficiency of both materials significantly decreased. This reduction is attributed to the formation of insulating iron carbonate precipitates (e.g., siderite) on the mZVI surface,

which could retard the electron transportation (Yin et al., 2012). Moreover, although HA imposed impact on the decontamination efficiency of both materials, Fe–N–C/ZVI was relatively less affected than mZVI (Fig. S14d), which could be attributed to the fact that the modification of the Fe–N–C catalytic layer could effectively alleviate the inactivation of reaction sites that are usually caused by the quick complexation of macromolecular HA with iron species (Zhang et al., 2009). Given the prevalence of NOM in natural environments, Fe–N–C/ZVI is anticipated to offer superior decontamination stability and efficiency under realistic conditions.

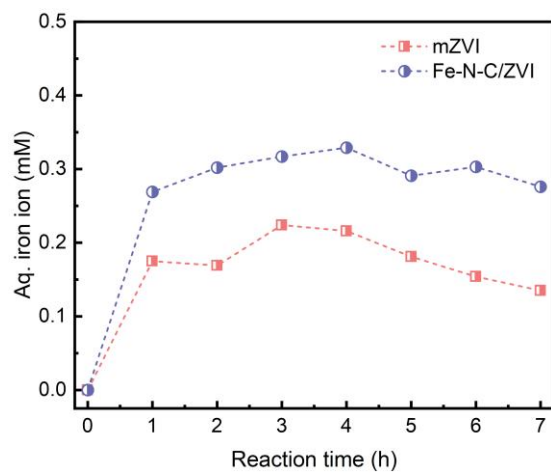


**Fig. S15.** Comparative evaluation of NB removal efficiency and kinetic behavior under varying pH conditions. Subfigure pairs (a-b) pH=3, (c-d) pH=5, (e-f) pH=7, and (g-h) pH=9 show temporal removal profiles (left panels) and corresponding kinetic fittings (right panels), respectively. ( $[NB]_0=100$  mg/L,  $T=298$  K).



**Fig.**

**S16.** The cyclic fixed-bed decontamination performances of Fe-N-C, mZVI, and Fe-N-C/ZVI toward (a) synthesized wastewater and (b) real wastewater collected from an industrial park (T=298 K, flow rate=6 BV/h).



**Fig. S17.** The iron ion concentration was detected in the effluent from continuous-flow degradation experiments. ( $[\text{NB}]_0=60 \text{ mg/L}$ ,  $T=298 \text{ K}$ , flow rate=6 BV/h).

**Table S1.** AC-impedance spectrum (EIS) fitting parameters for different catalysts.

Materials	$R_s$ ( $\Omega \cdot \text{cm}^2$ )	$R_{ct}$ ( $\Omega \cdot \text{cm}^2$ )	$W_o-R$ ( $\Omega \cdot \text{cm}^2$ )	<i>error</i> (%)
Fe-N-C	0.507	0.124	0.1415	4.81
NPCs	0.829	1.053	2.421	3.56
GR	1.4	2.24	1.861	1.78

**Table S2.** GC–MS analysis of intermediates in the electrocatalytic reduction of nitrobenzene using Fe–N–C electrode.

Compound	Chemical formula	Molecular weight	Mass of predominant ions in the fragmentation pattern	Retention time (min)
Nitrobenzene (NB)	$C_6H_5NO_2$	123.11	51, 60, 65, 77, 87, 93, 107, 115, 123 ( $M^+$ )	16.96
Nitrosobenzene (NSB)	$C_6H_5NO$	107.11	51, 60, 63, 66, 74, 77, 107 ( $M^+$ )	12.36
Phenyl hydroxylamine (PHA)	$C_6H_7NO$	109.13	39, 51, 65, 66, 77, 80, 92, 93, 109 ( $M^+$ )	11.19
Aniline (AN)	$C_6H_5NH_2$	93.13	52, 55, 60, 63, 66, 74, 78, 90, 93 ( $M^+$ )	7.63
Azobenzene (AB)	$C_{12}H_{10}N_2$	182.22	51, 57, 63, 71, 77, 91, 105, 115, 128, 139, 152, 182 ( $M^+$ )	25.03

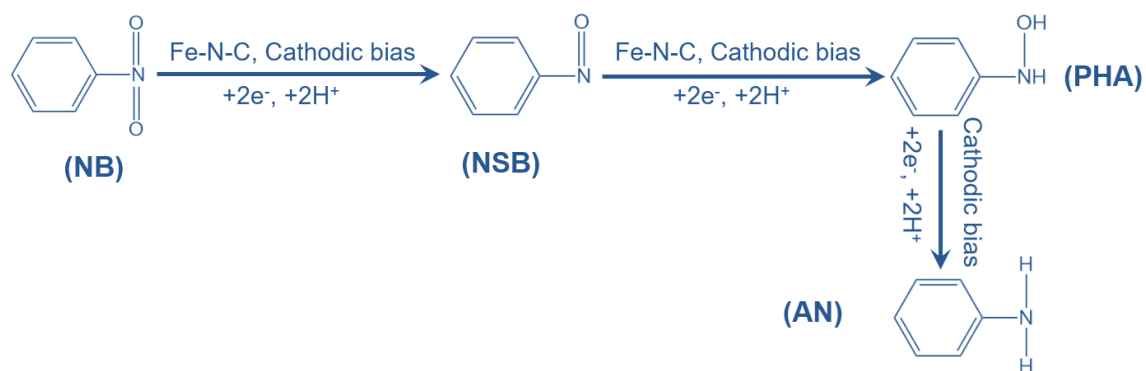
**Table S3.** Corrosion currents and potentials for different materials obtained from polarization curves via Tafel extrapolation.

Materials	$E_{corr}$ (V, vs SCE)	$i_{corr}$ ( $\mu\text{A}/\text{cm}^2$ )
mZVI	-0.413	5.724
ACs/ZVI	-0.501	5.910
NPCs/ZVI	-0.562	7.036
Fe-N-C/ZVI	-0.621	10.393

**Table S4.** Comparison of chemical and electrochemical systems used for nitrobenzene degradation.

Catalyst	Condition	pH	Concentration of NB	NB removal	Reduction rate	Ref
Nitrogen-doped porous carbon	-0.8 V vs. SCE	5.0	100 mg/L	94% (180 min)	0.0167 min <sup>-1</sup>	(Zhao et al., 2020)
Graphite	-0.8 V vs. SCE	5.0	100 mg/L	47% (180 min)	0.004 min <sup>-1</sup>	(Zhao et al., 2020)
Fe <sup>0</sup> /PVA microspheres	0.2 g/L	6.0	1 M	89% (200 min)	0.0027 min <sup>-1</sup>	(Bai et al., 2009)
C-dots@ZVI	2 g/L	5.0	50 mg/L	100% (120 min)	0.0596 min <sup>-1</sup>	(Fan et al., 2021)
mZVI particle (1 mm)	20 g/L	3~9	0.066 mM	87.5% (240 min)	0.0185~0.0038 min <sup>-1</sup>	(Dong et al., 2010)
TiO <sub>2</sub> NTAs	-1.2 V vs. SCE	5.2	300 mg/L	95% (120 min)	0.0242 min <sup>-1</sup>	(Ahmadi and Wu, 2019)
CNT-modified PGE	-1.2 V vs. SCE	5.0	1 mM	95% (50 min)	/	(Li et al., 2007)
Electrolysis/biochar	5V (1 g/L)	5.8	100 mg/L	99.2% (120 min)	0.0181 min <sup>-1</sup>	(Liu et al., 2020)
TiO <sub>2-x</sub> single crystals	-1.2 V vs. SCE	6.0	10 mg/L	90% (60 min)	0.0375 min <sup>-1</sup>	(Liu et al., 2016)
Nitrogen-doped diamond (NDD)	-0.9 V vs. SCE	6.0	100 mg/L	73.6% (150 min)	0.0155 min <sup>-1</sup>	(Zhang et al., 2014)
Electrolysis/redox mediator	0.04 V vs. RHE (1 mM PW12)	1.0	10 mM	100% (180 min)	0.0339 min <sup>-1</sup>	(Liu et al., 2024)
Zn/Fe-modified biochar	2 V (1 g/L)	5.6	100 mg/L	94% (120 min)	0.0147 min <sup>-1</sup>	(Liu et al., 2021)
Cu <sub>2</sub> O NPs/BiO <sub>2-x</sub> NSs	10 mA/cm <sup>2</sup>	7.0	100 mg/L	90% (30 min)	/	(Chen et al., 2022)
(FeCoNiCuZn)O/graphite	-1.7 V vs. Ag/AgCl	6.0	31 mg/L	90% (240 min)	0.0156 min <sup>-1</sup>	(Márquez et al., 2023)
Fe-N-C	-1.0 V vs. Ag/AgCl	3~9	100 mg/L	95% (120 min)	0.0377~0.0232 min <sup>-1</sup>	This work
Fe-N-C/mZVI	1 g/L	3~9	100 mg/L	100% (60 min)	0.021~0.009 min <sup>-1</sup>	This work

**Scheme S1.** Proposed pathway for NB electrochemical reduction by Fe–N–C electrode.



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